Nitrogen Isotope Compositions of Synthetic Fertilizer, Raw Livestock Manure Slurry, and Composted Livestock Manure

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To investigate the difference in N isotope ratio $({}^{15}N/{}^{14}N)$, expressed as $\delta^{15}N)$ among N sources (synthetic fertilizer, livestock manure, and manure compost), eight synthetic fertilizer, four livestock manure, and thirty-seven compost samples were collected and analyzed for $\delta^{15}N$. The mean $\delta^{15}N$ values of N sources were $-1.5 \pm 0.5\%_0$ (range: -3.9 to $+0.5\%_0$) for synthetic fertilizer, $+6.3 \pm 0.4\%_0$ (+5.3 to $+7.2\%_0$) for manure, and $+16.0 \pm 0.4\%_0$ (+9.3 to $+20.9\%_0$) for compost. The lower $\delta^{15}N$ of synthetic fertilizer was attributed to its N source, atmospheric N₂ of which $\delta^{15}N$ is 0‰. Meanwhile, more ${}^{15}N$ -enrichment of compost than manure was assumed to be resulted from N isotopic fractionation (faster loss of ${}^{14}N$ -bearing compound than ${}^{15}N$) associated with N loss particularly via NH₃ volatilization during composting. Therefore, our study shows that $\delta^{15}N$ values could successfully serve in discriminating two major N sources (synthetic fertilizer and compost) in agricultural system.

Key words: Livestock manure compost, Nitrogen isotope ratio, Nitrogen isotopic fractionation, Nitrogen source, Synthetic fertilizer

Introduction

The N isotope ratio (${}^{15}N/{}^{14}N$, expressed as $\delta^{15}N$) of nitrogenous compounds have been intensively used in tracing the fate of synthetic fertilizer in soil (Broadbent et al., 1980; Karamanos and Rennie, 1980; Karamanos and Rennie, 1981a, b), plant (Kohl et al., 1973; Meints et al., 1975; Shearer and Legg, 1975; Broadbent et al., 1980), and groundwater (Kreitler, 1979; Flipse and Bonner, 1985; Komor and Anderson, 1993; Feast et al., 1998). These works are possible due to considerably lower $\delta^{15}N$ values of synthetic fertilizer than those of natural soil N. Since synthetic N fertilizer is produced through the Haber-Bosch process, which uses atmospheric N₂ (the $\delta^{15}N$ is 0‰) as the reactant, the $\delta^{15}N$ values of synthetic fertilizer are close to 0‰. For groundwater, contribution of livestock feedlot to NO₃⁻ contamination has been estimated through

 δ^{15} N investigation as δ^{15} N values of livestock manure N are higher than those of soil-N via N isotope fractionation (faster reaction of ¹⁴N-bearing compounds than ¹⁵N-bearing N compounds) during N loss such as ammonia volatilization (Gormly and Spalding, 1979; Heaton, 1984).

More recently, the use of the $\delta^{15}N$ of plant tissues as a potential technique to identify the use of synthetic fertilizer in cultivating has been extensively studied in an effort to develop a technique for certification of organic produce that was grown without application of synthetic fertilizers (e.g., Choi et al., 2002; 2003; 2006; Bateman et al., 2005). The principle of this technique is quite similar to that used in identifying NO3⁻ contamination sources of groundwater (Choi et al., 2007); the δ^{15} N of synthetic N fertilizer manufactured from atmospheric N₂ using the Haber-Bosch process is usually lower than that of organic source of N fertilizers (e.g. livestock manure and its compost) that can be used in organic farming (Choi et al., 2003). However, up to now, there is limited information on the distribution of δ^{15} N amongst synthetic fertilizer, manure. and compost that is basic data for utilization of $\delta^{15}N$ in

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identification of N sources (Bateman and Kelly, 2007). In this study, δ^{15} N values of different N sources (synthetic fertilizer, manure, and compost) were investigated to confirm their differences in N isotopic compositions.

Materials and Methods

Sample collection Urea [(NH₂)₂CO], ammonium sulfate [(NH₄)₂SO₄], and six types of compound fertilizers (N-P₂O₅-K₂O ratio: 11-12-10, 12-8-12, 13-0-13, 15-15-15, 16-0-12, and 21-17-17) were collected. Four livestock manure samples (two pig, one cattle, and one poultry) were taken from the livestock feedlots around Gyonggi and Chonnam Provinces. Thirty-seven samples of commercially available composts were collected from compost manufacturing factories. Manure type of the compost samples are provided with their N concentration and δ^{15} N in Table 1.

Chemical analyses Synthetic fertilizers were digested by permanganate-reduced iron modification of Kjeldahl method to include nitrate and nitrite and steam-distilled with 10 N NaOH (Bremner, 1996). The liberated NH₃was trapped in 0.01 N H₂SO₄ and titrated with 0.01 N NaOH to determine the amount of NH₄⁺. The pH of the titrated solution was adjusted to 2~3 using 0.1 N H₂SO₄ or 0.1 N NaOH, and then water in the solution was evaporated to dryness under an infra-red lamp (Hauck, 1982). Nitrogen gas was prepared through the reaction of (NH₄)₂SO₄ and alkaline LiOBr under vacuum (Hauck, 1982). The δ^{15} N values of N₂were determined using a stable isotope ratio mass spectrometer of dual inlet type (VG OPTIMA, Micromass, UK).

The manure slurry samples were passed through a Whatman #42 filter paper under suction to remove coarse solid particles. The samples were processed following the same procedure as synthetic fertilizer analysis.

Composts samples were dried with a freeze drier to avoid δ^{15} N changes via N isotopic fractionation associated with NH₃ volatilization during drying and the dried samples were ground to a fine powder with a ball mill. The N concentrations and the corresponding δ^{15} N values of the compost samples were determined using a continuous-flow stable isotope ratio mass spectrometer linked with a CN analyzer (IsoPrime-EA, Micromass, UK).

Nitrogen isotope compositions were calculated as

ID	Manure type	N concentration (%)	δ ¹⁵ N (‰)	ID	Manure type	N concentration (%)	δ ¹⁵ N (‰)
1	Pig	1.44	+16.7	21	Pig, Cattle	1.46	+15.8
2	Poultry, Pig	2.10	+17.4	22	Pig, Cattle	1.45	+19.8
3	Pig	1.57	+17.7	23	Pig	1.14	+17.9
4	Pig	1.09	+18.2	24	Pig	1.20	+19.2
5	Pig	1.88	+16.8	25	Pig	1.23	+20.9
6	Poultry, Pig	2.45	+16.6	26	Pig, Cattle	1.60	+16.6
7	Pig	1.35	+15.4	27	Pig, Cattle	1.71	+15.2
8	Poultry, Pig	2.69	+18.8	28	Poultry, Pig	3.20	+12.0
9	Pig, Cattle	1.91	+19.4	29	Poultry, Pig	3.50	+9.3
10	Pig	1.88	+17.3	30	Pig, Cattle	1.60	+13.7
11	Pig, Cattle	2.09	+17.2	31	Pig, Cattle	1.80	+13.8
12	Pig, Cattle	2.15	+14.8	32	Pig	1.50	+16.6
13	Pig	1.66	+14.8	33	Pig, Cattle	1.90	+15.8
14	Pig	1.51	+16.0	34	Pig, Cattle	2.30	+14.6
15	Poultry, Pig	2.41	+17.2	35	Pig	1.40	+13.8
16	Pig	1.04	+15.3	36	Pig, Cattle	1.80	+17.2
17	Pig, Cattle	1.51	+14.5	37	Pig, Cattle	1.90	+15.1
18	Pig	1.19	+12.3				
19	Pig	1.06	+18.9				
20	Pig, Cattle	1.89	+10.0				

Table 1. Manure type, N concentration, and $\delta^{15}N$ of livestock manure composts (n=37).

$\delta^{15}N$ (‰) = [(R_{sample}/R_{standard})-1]×1000

where R_{sample} and $R_{standard}$ are the ratio of ${}^{15}N/{}^{14}N$ for sample and standard (atmospheric N₂), respectively. The accuracy and reproducibility of measurement checked with a reference material (RM 8548: IAEA-N2) from International Atomic Energy Agency were better than 0.4 and 0.2‰, respectively. All the chemical analyses were triplicated for each sample.

Statistical analysis Statistical analysis was carried out with PASW Statistics (SPSS Inc. Chicago, IL). We report the range, frequency distribution, and means (with standard errors) of δ^{15} N values without statistically based comparison due to the inherent limitation of the study design (pseudo-replicated samples). Pearson correlation analysis was used to examine the relationship between concentration and δ^{15} N of compost N.

Results and Discussion

The mean values of δ^{15} N of synthetic fertilizer, manure, and compost were -1.5 ± 0.5‰ (ranged from -3.9 to +0. 5‰), +6.3 ± 0.4‰ (from +5.3 to +7.2‰), and +16.0 ± 0. 4‰ (from +9.3 to +20.9‰), respectively. Therefore, the δ^{15} N of three sources were clearly separated each other without overlapping; synthetic fertilizers lower than +1. 0‰, manures between +5.0 and +8.0‰, and composts over +9.0‰ (Fig. 1). Around 50% of the 37 compost samples had δ^{15} N between +15.0 and +18.0‰ (Fig. 1).

The δ^{15} N values of synthetic fertilizers (-3.9 ~ +0.5‰) observed in this study were in good agreement with other studies. Overall, the δ^{15} N values for synthetic fertilizer are



Fig. 1. Histograms of δ^{15} N values of synthetic fertilizer (n=8), manure (n=4), and compost N (n=37). Arrows show the observed range of each N source.

between -8.0 and +6.2‰ with 90% ranging from -3.0% to +2.0‰ (Freyer and Aly, 1974; Kreitler, 1977; Mariotti and Letolle, 1977; Kreitler et al., 1978; Bateman and Kelly, 2007). Such a relatively low $\delta^{15}N$ of synthetic fertilizer reflects their N source; atmospheric N₂ which has a $\delta^{15}N$ value of 0‰, and the differences in $\delta^{15}N$ among fertilizers are generally considered a result of fractionation during subsequent processing of the N fixed through the Haber-Bosch process (Flipse and Bonner, 1985).

The δ^{15} N of cattle and pig manure slurry ranging from +5.3 to +7.2% suggests that the manure samples have been experienced with N loss (particularly via NH₃ volatilization) to some degree that causes N isotopic fractionation as original N excreted from animal is depleted in ¹⁵N, while N incorporated into protein retaining in animal body is enriched in ¹⁵N (Kreitler, 1975 Gormly and Spalding 1979). For examples, Kreitler (1975) reported that δ^{15} N of the original urea in the excreted N was -3.5‰. Gormly and Spalding (1979) have also shown that $\delta^{15}N$ of fresh animal manure was not different from $\delta^{15}N$ of soil- or fertilizer-derived N (<+5.0‰). Because of high urea-N concentration (accounts for up to 50% of total N) in livestock manure, volatilization of NH3 following urea hydrolysis is the predominant pathway of N loss from manure after excretion (Kerley and Jarvis, 1996). It has been shown that the ratio of ¹⁴NH₃ to ¹⁵NH₃ volatilization rate is between 1.020 and 1.029, indicating that volatilization of 14 NH₃ is faster than 15 NH₃ by 2.0 to 2.9% (adopted from Chang and Choi, 2009). Choi (2002) demonstrated ¹⁵N enrichment of manure slurry by NH₃volatilization through a laboratory incubation study with human, pig, and cattle manure; δ^{15} N of manure N increased from +11.8 to +16. 7‰ for human manure, +10.3 to +51.4‰ for pig manure, and +7.2 to +33.1‰ for cattle manure with decreasing NH_4^+ concentration in the manure samples.

Enrichment of ¹⁵N in N derived from manure can further progressed during composting as around 50% of N in the manure can be lost mainly via NH₃volatilization (Hao et al., 2004; Larney et al., 2006). In our study, the higher δ^{15} N of compost than raw manure could be an indicative of the ¹⁵N enrichment resulting from N isotopic fractionation associated with N loss (Fig. 1). Meanwhile, variation in δ^{15} N compost might be due to the different magnitude and extent of NH₃ volatilization among composts being affected by composting conditions and materials used. Kim et al. (2008) found that δ^{15} N of N in cattle manure composting mixture increased from +7.8‰



Fig. 2. Relationship between N concentration and corresponding δ^{15} N of livestock manure composts (n=37).

at the initiation of composting to +9.9‰ when rice hull was used as bedding material and from +11.4 to +14.3‰ when saw dust was mixed with the manure. They attributed such different pattern to contrasting characteristics of rice hull and saw dust in terms of their $\delta^{15}N$ and decomposability. The negative correlation between N concentration and the corresponding $\delta^{15}N$ of compost N is also notable (Fig. 2). This relationship suggests that $\delta^{15}N$ of compost increases with decreasing N concentration due to ^{15}N enrichment of compost via N loss.

Conclusions

The fact that $\delta^{15}N$ values of compost were higher than those of synthetic fertilizer suggests that application of compost to fields may result in different $\delta^{15}N$ of nitrogenous compounds in soil-plant-groundwater system, compared with synthetic fertilizer application. Therefore, this study shows the potential use of $\delta^{15}N$ technique in discriminating compost-derived N and fertilizer-derived N in agricultural system.

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화학비료, 가축분뇨 및 퇴비의 질소동위원소비

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화학비료, 가축분뇨 및 퇴비 등 주요 질소원의 질소동위원소비 (δ¹⁵N) 차이를 조사하기 위해 각각 8, 4, 37점의 시료를 채 취하여 δ¹⁵N을 분석하였다. 평균 δ¹⁵N 값은 화학비료가 -1.5 ± 0.5‰ (범위: -3.9~+0.5‰‰), 가축분뇨가 +6.3 ± 0.4‰ (+5.3~+7.2‰), 가축분퇴비가 +16.0 ± 0.4‰ (+9.3~+20.9‰)였다. 화학비료가 타 질소원에 비해 δ¹⁵N 값이 낮은 것은 화학비료 제조시 이용하는 질소원인 대기 N₂의 δ¹⁵N 값 (0‰)을 반영하기 때문이다. 반면, 가축분에 비해 퇴비의 δ¹⁵N 값 이 높은 것은 퇴비화 과정 중 일어나는 질소손실 (특히, 암모니아 휘산)과 관련된 질소동위원소분할효과 (¹⁴N의 손실속 도〉¹⁵N의 손실속도)에 의한 퇴비 중 ¹⁵N 농축에 의한 결과로 판단된다. 따라서, 본 연구는 δ¹⁵N 분석을 통해 현재 우리나라 농업 시스템에서 가장 널리 이용되고 있는 두 가지 질소원 (화학비료와 퇴비)을 구분할 수 있음을 보여준다.